

Marine phytoplankton community structure of the waters of the Argentine Islands (2019-2021)

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Abstract

The variability of phytoplankton community structure was analyzed in the waters of the Argentine Islands in late summer period of 2019, 2020 and 2021, as well as from February 2020 to February 2021. Biodiversity (121 taxa) and phytoplankton structure of the Argentine Islands waters corresponded to earlier studies. The late summer succession in 2019 and 2020 and 2021 was found typical for the waters of the Argentine Islands. Representatives of *Bacillariophyceae* completely dominated the studied community. Biomass variability was largely determined by large centric diatoms with low specific surface (S/W) values ranging from 70 to 300 m²·kg⁻¹. Higher values of phytoplankton development metrics and a more noticeable presence of haptophyte, cryptophyte and diatom groups with S/W values larger than 700 m²·kg⁻¹ differed 2019 from 2020 and 2021. Lower development metrics and simplified taxonomic structure of phytoplankton were recorded in 2020 – 2021. The decrease in the specific surface of the groups in the late summer periods of 2020 and 2021 was caused both by a decrease in the relative contributions of small-celled *Coccolithophyceae* and *Cryptophyceae* and by a general decrease in S/W of these taxa. Characteristic annual phytoplankton development patterns were revealed. The spring phytoplankton development was formed by high abundance of *Fragilariopsis* with a predominant contribution to the biomass of large centric diatoms. At the beginning of austral summer, the role of flagellates from the genera *Cryptomonas* (*Cryptophyta*) and *Pyramimonas* (*Chlorophyta*), as well as small flagellates, increased. The late austral summer period was characterized by the dominance of large centric diatoms with a significant presence of small flagellates and a periodic increase in the contribution of *Fragilariopsis*. The spring-summer succession had common features with the development of phytoplankton in the more northern regions of the Western Antarctic Peninsula (Palmer station), while the late summer phase – with the more southern Rothera station.

Key words: phytoplankton, Western Antarctic Peninsula (WAP), Argentine Islands, succession, monitoring, biomass, specific surface.

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Introduction

Antarctic coastal areas are characterized by high primary productivity (Arrigo *et al.* 2017). In the waters of the Western Antarctic Peninsula (WAP), it is close to other areas of the continental shelf of Antarctica and is about $182 \text{ gC}\cdot\text{m}^{-2}$ per year (Arrigo *et al.* 2008), which is about four times lower than in the most productive coastal regions of the World Ocean (Vernet and Smith 2007, Nardelli *et al.* 2023). This difference is naturally determined by the temperature, light and ice cover, that limit the level of WAP phytoplankton development, especially in the autumn-winter period. The main phytoplankton production is formed during spring-summer development, when chlorophyll concentrations in WAP waters can exceed $20 \mu\text{g}\cdot\text{l}^{-1}$ (Vernet *et al.* 2008, Ducklow *et al.* 2013, Kim *et al.* 2018). Phytoplankton productivity and structure affects the development of all subsequent components of the Antarctic food chain (Moline *et al.* 2004, Saba *et al.* 2014, van Leeuwe *et al.* 2020). In addition, microalgae play an important role in the response of the Antarctic ecosystem to climate change through the absorption of CO_2 (Legge *et al.* 2017) and the production of climate-active gases (Stefels *et al.* 2018, Webb *et al.* 2019, Hughes *et al.* 2009).

The WAP has experienced significant and rapid climatic changes in recent decades (Meredith and King 2005, Stammerjohn *et al.* 2008). During the second half of the twentieth century, WAP warming rates were the largest in the Southern Hemisphere, leading to an increase in ocean surface temperature of more than 1°C (Meredith and King 2005, Turner *et al.* 2005). These processes have continued after a certain period of stabilization (Turner *et al.* 2016). As of 2016, 90% of sea glaciers have retreated, and the annual sea ice season has shortened by >92 days since 1979 (Stammerjohn *et al.* 2012, Cook *et al.* 2016). Further warming in Western Ant-

arctica is predicted during the 21st century (Meijers 2014).

Biological systems of Antarctic coastal areas are extremely sensitive to climate change due to the adaptation of marine organisms to a limited range of environmental conditions at the level of diversity, physiology, and involvement in marine biogeochemical cycles (Steiner and Stefels 2017). The close relationship between environmental conditions and phytoplankton components (community structure) suggests that continued warming in the WAP will affect the seasonal dynamics of phytoplankton and subsequently the seasonal dynamics of the food web (Nardelli *et al.* 2023). It is expected that ocean warming, sea ice and glacier retreat, causing increased spatial distribution of low-salinity surface waters, will lead to higher abundance of small-celled phytoplankton along the WAP (Moline *et al.* 2004). This will have important implications for food web structure and for the efficiency of trophic energy transfer (Sailley *et al.* 2013). So far, there is a trend towards a decrease in the summer phytoplankton biomass and in the size of the cells forming it in the northern WAP (Montes-Hugo *et al.* 2008, 2009; Nardelli *et al.* 2023). It is assumed that such morphological rearrangement is caused by an increase in the number of small-celled cryptophytes in coastal regions (Moline *et al.* 2004, Mendes *et al.* 2013, 2018; Schofield *et al.* 2017), associated with broader adaptation of cryptophytes to light conditions (Kaňha *et al.* 2012, Nardelli *et al.* 2023). In addition to cryptophytes, the abundance of other flagellates, such as haptophytes (Arrigo *et al.* 2017, Stefels *et al.* 2018), prasinophytes and chlorophytes (van Leeuwe *et al.* 2020), is predicted to increase. Expected near-future conditions with a more dynamic ice cover may result in increased primary production of particular species, as well as in overall algal biomass, with more scattered

wind-mixed summer blooms (van Leeuwe et al. 2020).

Phytoplankton responses to climate change are difficult to model and predict, as the studies on its structural and functional reorganization in response to the environmental factors are scarce. Meanwhile, the model predictions available to date cover large planetary scales (Henson et al. 2016). Decades of observations are required for the reliable regional modelling of the climate change impact, which is of utmost importance for the functioning of Antarctic ecosystems (van Leeuwe et al. 2020). Thus, phytoplankton monitoring is essential for the analysis of complex ecosystem seasonal responses and general trends (Deppeler and Davidson 2017, Henley et al. 2019). Palmer (Anvers Island) and Rothera stations (Adelaide Island) are the main locations of the WAP, where

systematic studies of phytoplankton are carried out (van Leeuwe et al. 2020, Nardelli et al. 2023). Vernadsky station (Argentine Islands), where phytoplankton monitoring studies have also been conducted since 2002 (Seregin et al. 2005, Kuzmenko and Ignatiev 2007), is located between them, 50 km to the southwest of the Palmer station and 200 km to the northeast of the Rothera station. The fastest warming rates for WAP were recorded in the waters of the Argentine Islands (Khrystiuk et al. 2022), which make it a particularly relevant polygon for monitoring of phytoplankton, as the first responders to climate change. Thus, the purpose of the present study is to analyze the variability of phytoplankton community structure in the Argentine Islands' waters in 2019-2021 and to define the trends of its dynamics.

Material and Methods

Study site

Research was conducted during the summer seasons of the 24th, 25th, and 26th Ukrainian Antarctic expeditions, from February to April 2019, 2020, and 2021. Samples were also collected during the period of April 2020 – February 2021. In 2019 and 2020, samples were taken at 6 stations in the waters of the Argentine Islands archipelago (Fig. 1) (Supplementary, Table 1). The analysis of the results

obtained in 2019-2020 showed a stable lack of statistically significant difference between the data obtained from some stations. Thus, the marine monitoring scheme was optimized in 2021 and sampling was carried out at three stations (1, 3 and 5) during the seasonal expedition, and at station 3 during the annual monitoring (April 2020 – February 2021) (Fig. 1.).

Sampling

Sampling was performed weekly during the seasonal expeditions of 2019, 2020 and 2021 and once every two weeks during the 2020-2021 annual monitoring. The vertical profile of salinity, temperature, fluorescence, and oxygen concentration was recorded with the hydrochemical probe (Rbr Concerto, Canada) at each sampling event.

Water samples for phytoplankton quantitative analysis were taken with a 5-liter Niskin bottle at the surface (1 m) and in the deep fluorescence maximum (DCM) layer. Qualitative phytoplankton samples were taken at each of the monitoring stations using a plankton net with a mesh size of 5 μ m. Quantitative analysis was per-

formed for 177 phytoplankton samples, while 258 samples were analyzed for the concentration of chlorophyll *a* and pheophytin (for the method, *see below*).

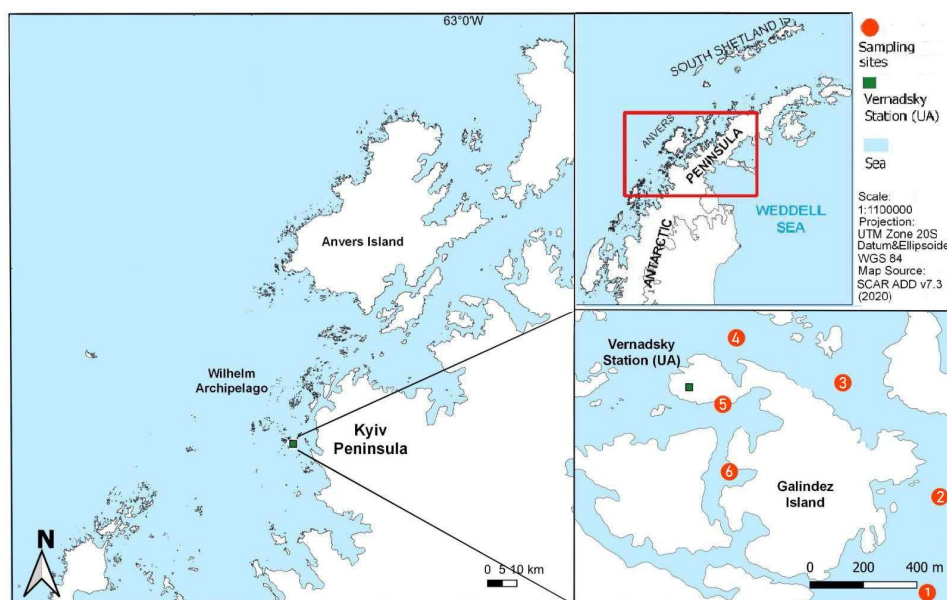


Fig. 1. Marine monitoring stations in the waters of the Argentine Islands archipelago, WAP.

Sample preparation and counting

Quantitative phytoplankton samples with a volume of up to 5 l were concentrated by reverse filtration through nitrocellulose filters with a pore diameter of 1.02 μm (Millipore) to a volume of 30–50 ml. Concentrated samples were fixed with a formaldehyde acidified with acetic acid to a final concentration of 4% formalin in the sample. If necessary, the samples were concentrated again by slow decantation to a volume of 10–15 ml before counting phytoplankton taxa. Quantitative and taxonomic processing of phytoplankton samples was carried out under a Zeiss Primo Star light microscope with magnifications of 400 – 1000. Cell counting was carried out in two replicates in a Najotta chamber with a volume of 0.05 ml. A Sedgwick-Rafter chamber with a volume of 1 ml was used for additional count-

ing of large cells. Wet biomass was calculated by geometric similarity of the cells' shape to the corresponding geometric shapes and assuming that the density of cells is equal to 1. Metrics of surface area (S), surface index (IS) and specific surface (S/W) for different levels of phytoplankton organization were calculated based on Minicheva *et al.* (2003). The taxonomy was aligned with The World Register of Marine Species (WoRMS).

Chlorophyll *a* samples were collected by vacuum filtration (Combisart.jet Sartorius) through nitrocellulose filters with a pore diameter of 0.45 μm (Millipore) in duplicate for each station, followed by treatment of the filters with hot ethanol (90%, 78°C) for pigments' extraction and subsequent centrifugation at 3000 rpm. Pigments' concentration was measured us-

ing Hach Lange DR3900 spectrophotometer. A solution of 0.3 ml hydrochloric

acid and 100 ml ethanol was used to acidify the samples during the measurement.

Data analysis

The data was normalized via log-transformation. Fisher's least significant difference (LSD) with a 95.0% confidence level was used to detect a statistically

significant difference between the means. Prism 10.1.1 (GraphPad Software) was used to create figures.

Results

Taxonomic composition and structure of phytoplankton

Representatives of 121 species and infraspecific taxa of unicellular algae, belonging to 12 taxonomic classes, were found in the Argentine Islands waters (Ta-

ble 1; Fig. 2). Additionally, 30 taxa were encountered sporadically. They require further identification, which in most cases involves using molecular biology methods.

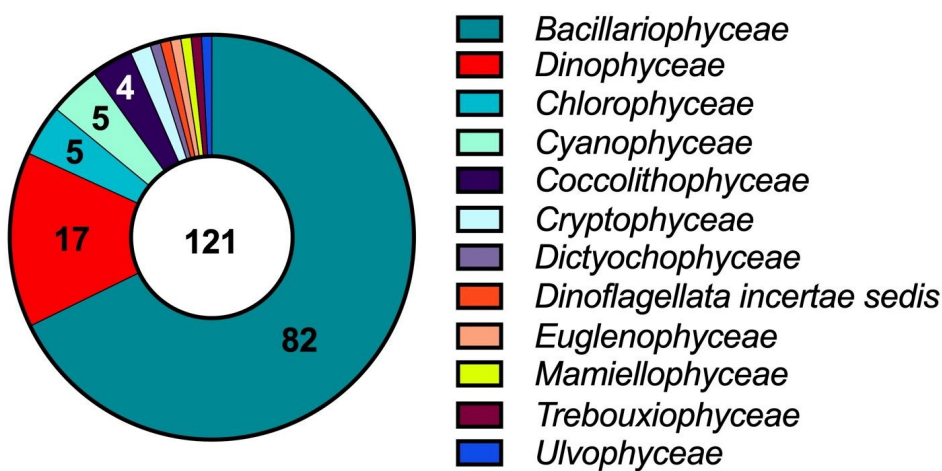


Fig. 2. Taxonomic distribution of the Argentine Islands phytoplankton at the class level based on the species number (February-April 2019, 2020 and 2021).

The highest species richness was found for the *Bacillariophyceae* class, which covered 68% of unicellular algae taxa (Fig. 2). *Dinophyceae* made up 14% of the total species number, and representatives

of *Chlorophyceae* and *Cyanophyceae* – 4% each. The other taxonomic groups' contribution to the phytoplankton diversity did not exceed 3% (Fig. 2).

Species	2019	2020	2021	$N_{\text{total}} \cdot 10^6 \cdot \text{m}^{-3}$	$B_{\text{total}} \cdot \text{mg} \cdot \text{m}^{-3}$
Bacillariophyceae					
<i>Achnanthes antarctica</i> Perag., 1921	+			1.73	5.68
<i>Achnanthes brevipes</i> C.Agardh, 1824	+	+		2.54	9.19
<i>Achnanthes longipes</i> C.Agardh, 1824			+	0.01	0.07
<i>Actinocyclus octonarius</i> var. <i>octonarius</i> Ehrenberg, 1837			+	0.01	0.06
<i>Amphiprora kjellmanii</i> var. <i>kjellmanii</i> Cleve in Cleve & Grunow, 1880	+	+	+	0.75	36.34
<i>Amphiprora paludosa</i> f. <i>hyperborea</i> Grunow in Cleve & Grunow, 1880	+			0.48	4.26
<i>Amphora ovalis</i> (Kützing) Kützing, 1844	+			0.98	11.66
<i>Amphora veneta</i> f. <i>veneta</i> Kützing, 1844	+	+	+	1.25	3.70
<i>Berkeleya antarctica</i> Grunow 1880	+	+	+	1.07	0.60
<i>Cerataulina bergonii</i> Ostensfeld, 1903		+		0.11	1.91
<i>Ceratoneis closterium</i> Ehrenberg, 1839	+	+	+	97.25	12.19
<i>Ceratoneis fasciola</i> Ehrenberg, 1839			+	0.00	0.06
<i>Chaetoceros compressus</i> Lauder, 1864	+	+	+	714.35	448.70
<i>Chaetoceros neogracilis</i> VanLand., 1968		+	+	0.31	0.15
<i>Chaetoceros socialis</i> H.S.Lauder, 1864		+		1.25	0.59
<i>Chaetoceros</i> sp.	+	+	+	5.46	10.08
<i>Chaetoceros tortissimus</i> Gran, 1900		+		0.70	0.52
<i>Charcotiella irregularis</i> (Perag.) S.Blanco & C.E.Wetzel, 2016	+			0.06	19.55
<i>Cocconeis adeliae</i> E. Manguin, 1960	+		+	0.50	38.49
<i>Cocconeis costata</i> var. <i>antarctica</i> Manguin, 1960	+	+	+	12.91	120.27
<i>Cocconeis infirmata</i> Manguin, 1957	+			0.65	4.75
<i>Cocconeis</i> sp.	+	+	+	0.10	4.18
<i>Corethron criophilum</i> Castracane, 1886	+	+	+	176.74	8890.13
<i>Corethron inerme</i> Karsten, 1905	+	+	+	7.18	2142.0
<i>Coscinodiscus oculus-iridis</i> (Ehrenberg) Ehrenberg, 1840	+			0.06	43.81
<i>Coscinodiscus radiatus</i> Ehrenberg, 1840		+		0.51	50.62
<i>Coscinodiscus</i> sp.	+	+	+	0.66	58.48
<i>Cyclotella</i> sp.	+			16.50	7.99
<i>Diatoma rhombica</i> E.O'Meara, 1877	+			3.92	2.48
<i>Diploneis elfvingiana</i> var. <i>latefurcata</i> C.W.Fontell, 1917			+	0.25	4.80
<i>Fragilaria nana</i> Steemann Nielsen, 1935		+		44.17	1.63
<i>Fragilariopsis curta</i> (Van Heurck) Hustedt, 1958		+		498.32	40.10
<i>Fragilariopsis cylindrus</i> (Grunow) Krieger, 1954	+	+		28990.13	570.19
<i>Fragilariopsis</i> sp.	+	+	+	89.80	102.06
<i>Grammatophora arcuata</i> Ehrenberg, 1853	+			0.84	2.97
<i>Haslea</i> sp.	+	+		0.33	10.48
<i>Licmophora abbreviata</i> C.Agardh, 1831	+	+	+	3.35	43.04
<i>Licmophora antarctica</i> Carlson, 1913	+	+	+	7.07	281.35

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<i>Licmophora ehrenbergii</i> (Kützing) Grunow, 1867	+	+		2.94	49.53
<i>Licmophora</i> sp.	+	+	+	18.74	17.01
<i>Melosira antarctica</i> Van Heurck, 1909		+		17.45	6.91
<i>Melosira</i> sp.		+	+	0.70	0.80
<i>Navicula challengerii</i> Grunow, 1880	+		+	0.16	5.74
<i>Navicula cendronii</i> Manguin, 1960	+	+		9.08	14.82
<i>Navicula insuta</i> Manguin, 1960	+	+		2.88	0.49
<i>Navicula molesta</i> Mann, 1925	+			0.38	0.16
<i>Navicula</i> sp.	+	+	+	180.20	174.74
<i>Neomoelleria antarctica</i> (Castrac.) S.Blanco & C.E.Wetzel, 2016	+			1.12	19.98
<i>Nitzschia adeliae</i> Manguin, 1960	+		+	1.35	1.41
<i>Nitzschia capitellata</i> Hustedt, 1922	+	+	+	1.15	0.50
<i>Nitzschia longissima</i> (Brébisson) Ralfs, 1861	+			1.27	0.21
<i>Nitzschia sicula</i> (Castracane) Hustedt	+			2.59	0.87
<i>Nitzschia</i> sp.	+	+	+	4.00	4.60
<i>Nitzschia tenuirostris</i> Mer.		+		0.07	0.01
<i>Odontella aurita</i> (Lyngbye) C.Agardh, 1832	+	+	+	3.32	905.36
<i>Odontella</i> sp.	+	+	+	7.41	408.16
<i>Pinnularia directa</i> var. <i>directa</i> W. Smith, 1853	+			14.34	39.13
<i>Pinnularia</i> sp.	+	+		0.13	12.86
<i>Pleurosigma clevei</i> Grunow, 1880	+	+	+	2.90	128.61
<i>Pleurosigma</i> sp.	+	+	+	1.61	65.31
<i>Podosira vanheurckii</i> var. <i>vanheurckii</i> M. Peragallo, 1921	+			0.42	9.21
<i>Porosira glacialis</i> (Grunow) Jörgensen, 1905		+	+	4.56	88.08
<i>Proboscia alata</i> (Brightwell) Sundström, 1986	+			1.96	29.20
<i>Proboscia</i> sp.	+			1.12	15.49
<i>Pseudo-nitzschia delicatissima</i> (Cleve) Heiden, 1928	+	+	+	43.48	19.41
<i>Pseudo-nitzschia seriata</i> (Cleve) H.Peragallo, 1899	+	+	+	79.86	27.64
<i>Rhaphoneis fasciolata</i> var. <i>fasciolata</i> Ehrenberg, 1844	+		+	0.12	5.72
<i>Rhizosolenia imbricata</i> Brightwell, 1858	+			0.79	7.49
<i>Rhizosolenia truncata</i> G.Karsten, 1905	+	+		8.91	180.89
<i>Synedra adeliae</i> Manguin, 1957	+	+	+	4.36	12.09
<i>Synedra camtschatica</i> var. <i>antarctica</i> Manguin, 1960	+	+	+	4.26	29.02
<i>Synedra</i> sp.	+	+	+	2.94	1.41
<i>Thalassiosira australis</i> M.Peragallo, 1921			+	0.09	2.88
<i>Terebraria kerguelensis</i> E.O'Meara, 1876 †	+	+	+	43.04	56.82
<i>Thalassiosira antarctica</i> Comber, 1896	+	+	+	72.98	286.03
<i>Thalassiosira oliveriana</i> (O'Meara) I.V.Makarova & V.A.Nikolajev, 1983	+	+		0.60	38.39
<i>Thalassiosira rotula</i> Meunier, 1910	+	+	+	30.61	23.32
<i>Thalassiosira</i> sp.	+	+	+	29.39	581.94
<i>Thalassiothrix antarctica</i> A.Schimper ex G.Karsten, 1905	+			0.27	7.25

<i>Trachyneis aspera</i> (Ehrenberg) Cleve, 1894			+	0.00	0.24
<i>Trigonium arcticum</i> (Brightwell) Cleve, 1868			+	0.00	3.62
<i>Tropidoneis fusiformis</i> E.Manguin, 1957	+			0.48	1.92
Chlorophyceae					
<i>Ankistrodesmus antarcticus</i> Kol & E.A.Flint, 1968		+	+	72.23	2.79
<i>Chlorophyceae</i> gen. sp.			+	1.61	5.68
<i>Chlorophyceae</i> gen. sp. 1		+	+	69.38	7.27
<i>Chlorophyceae</i> gen. sp. 2		+	+	30.87	11.80
<i>Monoraphidium</i> sp.	+	+		0.68	0.01
Coccolithophyceae					
<i>Acanthoica acanthos</i> J.Schiller, 1925	+	+		8.30	24.94
<i>Emiliania huxleyi</i> (Lohmann) W.W.Hay & H.P.Mohler, 1967	+			34.65	8.12
<i>Phaeocystis antarctica</i> Karsten, 1905	+			4147.41	971.72
<i>Rhabdosphaera hispida</i> Lohmann, 1912	+			0.68	1.49
Cryptophyceae					
<i>Cryptomonas</i> sp.	+			1231.26	128.68
<i>Hillea fusiformis</i> (J.Schiller) J.Schiller, 1925	+	+		1554.89	445.26
Cyanophyceae					
<i>Cyanophyceae</i> gen. sp.		+	+	0.68	0.53
<i>Cyanophyceae</i> gen. sp. 1		+		0.20	1.19
<i>Johannesbaptistia pellucida</i> (Dickie) W.R.Taylor & Drouet, 1938		+	+	0.06	0.68
<i>Nostoc</i> sp.		+		0.60	0.16
<i>Oscillatoria tenuis</i> C.Agardh ex Gomont, 1892		+		74.50	0.12
Dictyochophyceae					
<i>Octactis speculum</i> (Ehrenberg) F.H.Chang, J.M.Grieve & J.E.Sutherland, 2017	+	+	+	2.75	8.29
Dinoflagellata incertae sedis					
<i>Protodinium simplex</i> Lohmann, 1908	+			411.95	188.95
Dinophyceae					
<i>Amphidinium crassum</i> Lohmann, 1908	+		+	2.83	14.25
<i>Glenodinium paululum</i> Lindemann, 1928		+		1.27	3.36
<i>Gymnodinium agilisforme</i> Schiller, 1928	+			14.75	16.76
<i>Gymnodinium biconicum</i> J.Schiller, 1928	+			0.33	0.34
<i>Gymnodinium najadeum</i> J.Schiller, 1928	+			2.90	4.98
<i>Gymnodinium</i> sp.	+	+	+	28.44	55.77
<i>Gymnodinium wulffii</i> J.Schiller, 1933	+			35.36	18.41
<i>Gyrodinium lacryma</i> (Meunier) Kofoid & Swezy, 1921	+	+		20.66	143.97
<i>Gyrodinium pingue</i> (Schütt) Kofoid & Swezy, 1921	+			1.54	23.95
<i>Gyrodinium</i> sp.	+	+		0.65	0.54
<i>Kapelodinium vestifici</i> (Schütt) Boutrup, Moestrup & Daugbjerg, 2016	+	+		79.15	65.56
<i>Oxytoxum turbo</i> Kofoid, 1907			+	0.23	0.57
<i>Pronoctiluca spinifera</i> (Lohmann) Schiller, 1932	+			0.04	0.01

<i>Prorocentrum micans</i> Ehrenberg, 1834	+		0.44	0.46
<i>Prorocentrum</i> sp.	+	+	5.42	5.09
<i>Protoperdinium smithii</i> H.Doan-Nhu, L.Phan-Tan & L.Nguyen-Ngoc 2018	+	+	6.11	8.90
<i>Scrippsiella acuminata</i> (Ehrenberg)				
Kretschmann, Elbrächter, Zinssmeister, S.Soehner, Kirsch, Kusber & Gottschling, 2015	+		0.03	0.02
Euglenophyceae				
<i>Euglena</i> sp.		+	0.10	1.20
Mamiellophyceae				
<i>Mantoniella</i> sp.		+	0.30	0.00
Trebouxiophyceae				
<i>Closteriopsis longissima</i> (Lemmermann)	+	+	0.87	0.10
Lemmermann, 1899				
Ulvophyceae				
<i>Hormidiopsis crenulata</i> (Kützing) Heering, 1914	+	+	1.19	31.17

Table 1. Taxonomic list of phytoplankton in the waters of Argentine Islands, WAP (2019-2021).

Large-celled centric diatoms *Corethron criophilum* and *Corethron inerme* and the haptophyte *Phaeocystis antarctica* were dominant in terms of biomass (Table 1). Lower proportion of overall biomass was formed by representatives of *Odontella aurita*, *Thalassiosira* sp., *Fragilariopsis cylindrus* and *Chaetoceros compressus* (Table 1). The largest total abundance was found for small-celled diatoms *Fragilariopsis cylindrus*, *Fragilariopsis curta* and *Chaetoceros compressus*, cryptophytes *Cryptomonas* sp., *Hillea fusiformis* and *Phaeocystis antarctica*. Dinoflagellata *Protodinium simplex* reached significant abundances in 2019 (Table 1).

Representatives of the *Bacillariophyceae* have completely dominated in the phytoplankton of the Argentine Islands waters during the late summer periods of 2019, 2020 and 2021 (Fig. 3).

The lowest average contribution of diatoms to phytoplankton abundance (79%), biomass (79%) and surface area (93%) was observed in 2019. This is due to the significant development of *Coccolithophyceae*, accounting for up to

12% of the abundance and 10% of the biomass, and *Cryptophyceae*. The contribution of *Bacillariophyceae* to biomass and to the surface area increased to 97% and 98% respectively in the late summer period of 2020 and 2021 (Fig. 3). The representatives of *Coccolithophyceae* and *Cryptophyceae* were almost absent in the phytoplankton communities of 2020 and 2021. Instead, a moderate development of *Chlorophyceae* and undefined small-celled flagellate algae, with an average size of 12 μm and a volume of 470 μm^3 , was observed and reached its maximum during the period of diatom decrease (Fig. 3).

The development of the most massive diatom – *Corethron criophilum* was a major factor behind the dynamics of phytoplankton biomass in the waters of the Argentine Islands during the late summer period of 2019, 2020 and 2021 (Fig. 4). This species was present during the entire period of research and significantly contributed to the increase in the community biomass, which reached its maximum at the end of March 2019, while also peaking in the second decade of February 2020,

March 2020, and March 2021 (Figs. 3, 4). The specific surface of *Corethron criophilum* was $218 \text{ m}^2 \cdot \text{kg}^{-1}$, which indicated the low potential functional capacity of this diatom (Zotov 2008). Additionally, the other large-celled diatoms with similar S/W values formed the majority of phytoplankton biomass in the late summer period. These are *Corethron inerme* ($101 \text{ m}^2 \cdot$

kg^{-1}), *Odontella aurita* ($159 \text{ m}^2 \cdot \text{kg}^{-1}$), *Odontella sp.* ($167 \text{ m}^2 \cdot \text{kg}^{-1}$), *Thalassiosira sp.* ($232 \text{ m}^2 \cdot \text{kg}^{-1}$), *Porosira glacialis* ($238 \text{ m}^2 \cdot \text{kg}^{-1}$) (Fig. 4). Thus, during the late summer seasons of all three years, the biomass was mainly formed by a group of large centric diatoms with a low S/W (from 70 to $300 \text{ m}^2 \cdot \text{kg}^{-1}$).

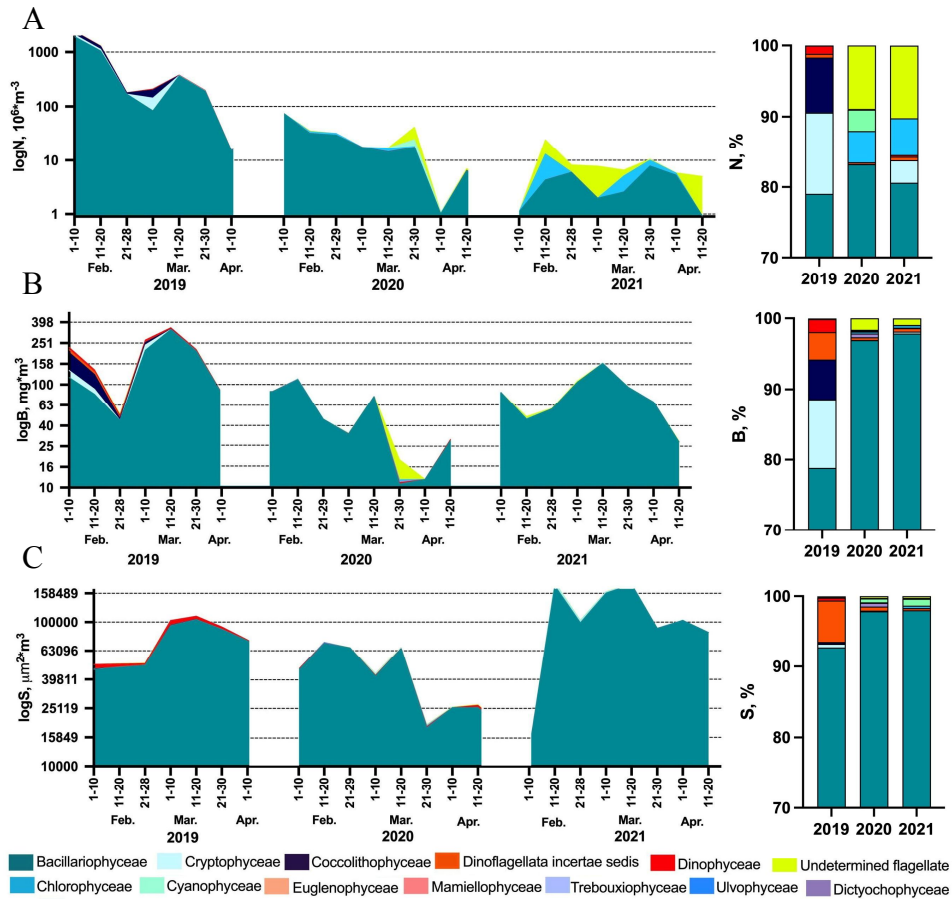


Fig. 3. Dynamics of the abundance (N), biomass (B) and surface area (S) of phytoplankton taxonomic groups in the waters of the Argentine Islands during the late summer period of 2019 – 2021 (see Supplementary, Table 3-5 for non-logarithmic values).

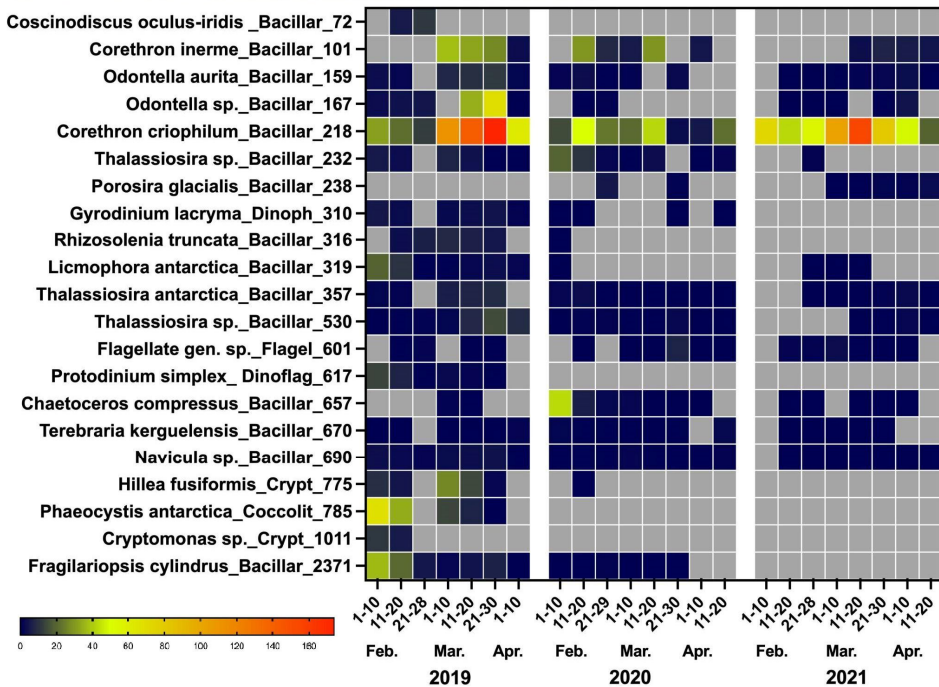


Fig. 4. Dynamics of the dominant phytoplankton species biomass ($\text{mg}\cdot\text{m}^{-3}$) in the waters of the Argentine Islands in late summer of 2019, 2020 and 2021 (X-axis: Decade – Month – Year; Y-axis: Species name - Taxonomic class - Specific surface value).

However, the nature of the late summer succession in 2019 and 2020 and 2021 was different. In 2019, the development of a group of small-celled species with specific surface values larger than $700 \text{ m}^2\cdot\text{kg}^{-1}$ was observed. At the beginning of February 2019, high abundance of *Phaeocystis antarctica* ($785 \text{ m}^2\cdot\text{kg}^{-1}$), cryptophyte *Hillea fusiformis* ($775 \text{ m}^2\cdot\text{kg}^{-1}$) and *Cryptomonas sp.* ($1011 \text{ m}^2\cdot\text{kg}^{-1}$) and small-cell pennate diatom *Fragilariopsis cylindrus* ($2371 \text{ m}^2\cdot\text{kg}^{-1}$) (Fig. 4) was detected. The gradual decrease of this group abundance during February indicates that seasonal studies may have revealed the end of a more massive bloom. These species were found in very low quantity in the following years. The group of species with S/W from 300 to $700 \text{ m}^2\cdot\text{kg}^{-1}$, including both diatoms and dinoflagellates, formed moderate biomass mainly during the late summer period of 2019 (Fig. 4). *Chaetoceros compressus*

diatom ($657 \text{ m}^2\cdot\text{kg}^{-1}$) was the only one reaching a significant biomass in early February 2020. A moderate development of undetermined small-celled flagellate *Flagellate gen. sp.* ($601 \text{ m}^2\cdot\text{kg}^{-1}$) was observed in March 2020, after the decline of *Corethron criophilum* abundance. Similar development of *Flagellate gen. sp.* happened in March 2021, between the two peaks of *Corethron criophilum* biomass (Figs. 3, 4).

Phytoplankton taxonomic structure, revealed during the period of April 2020 – February 2021, was characterized by the overwhelming dominance of *Corethron criophilum*. The contribution of large centric diatoms, such as *Porosira glacialis*, *Odontella sp.* and *Thalassiosira sp.*, to the community biomass increased in the end of summer, while the ratio of dinoflagellata *Peridinium sp.* – at its beginning.

Metrics of phytoplankton development

The interannual dynamics of the average phytoplankton abundance (N), biomass (B), surface area (S) and surface index (ISC) was synchronous for the late summer seasons of 2019, 2020 and 2021 (Fig. 5). These metrics were found to be significantly higher in 2019, than in 2020 and 2021 (Supplementary, Table 2), except for S values, which did not differ in 2019 and 2021 (Supplementary, Table 2).

2020 was characterized by the lowest values for phytoplankton development metrics (except for the abundance). Thus, the late summer seasons of three consecutive years were characterized by a decrease of the phytoplankton quantitative development metrics in 2020, and their increase in 2021 to a level that did not exceed the values of 2019.

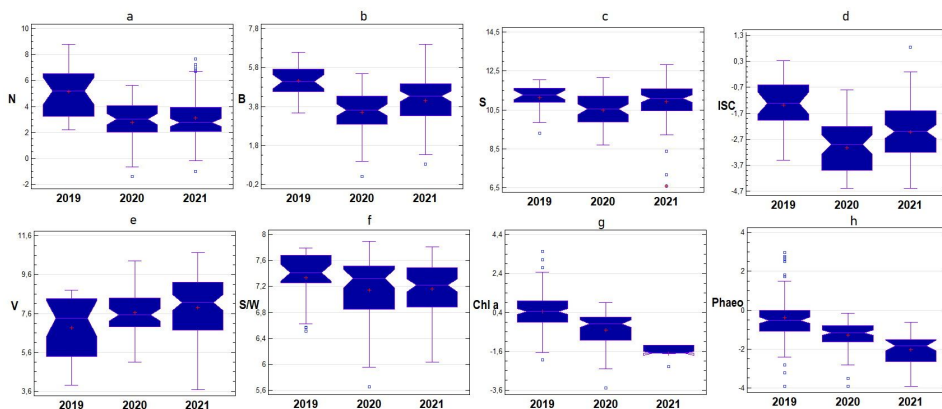


Fig. 5. Interannual variability of abundance (N, $10^6 \cdot \text{m}^{-3}$), biomass (B, $\text{mg} \cdot \text{m}^{-3}$), surface area (S, $\text{m}^2 \cdot \text{m}^{-3}$), surface index (ISC, m^{-1}), average cell volume (V, m^3), specific surface (S/W, $\text{m}^2 \cdot \text{kg}^{-1}$), concentrations of chlorophyll *a* (Chl *a*, $\text{mg} \cdot \text{m}^{-3}$) and phaeophytin (Phaeo, $\text{mg} \cdot \text{m}^{-3}$) of phytoplankton communities in the waters of the Argentine Islands in late summer period 2019–2021 (see Supplementary, Table 2 for Fisher's least significant difference).

The late summer dynamics of phytoplankton development metrics varied in different years (Fig. 6). A common feature was the presence of two peaks of phytoplankton development from the beginning of February to the end of April (Fig. 6). The biomass dynamics was most synchronous (Fig. 6b). After the minimum, observed from February 10 to March 10, in different decades depending on the year, an increase that reached a maximum on March 10–20 was detected, followed by a decrease to the April minimum (Fig. 6b). The ISC, which is most closely related to the gross primary production of phyto-

plankton (Zotov 2016), also reached maximum on March 10 – 20 in 2019 and 2020 (Fig. 6d). In 2021, the seasonal maximum values of the ISC were observed on February 21–30, while the March maximum of this metric was insignificant (Fig. 6d).

The dynamics of phytoplankton development, observed during the research from February 2020 to April 2021, indicates an early onset of the development. In 2020, the maximum of all metrics was observed in February (Fig. 7). In March 2020, a gradual decline began, reaching a minimum in June, changing little during the Antarctic winter.

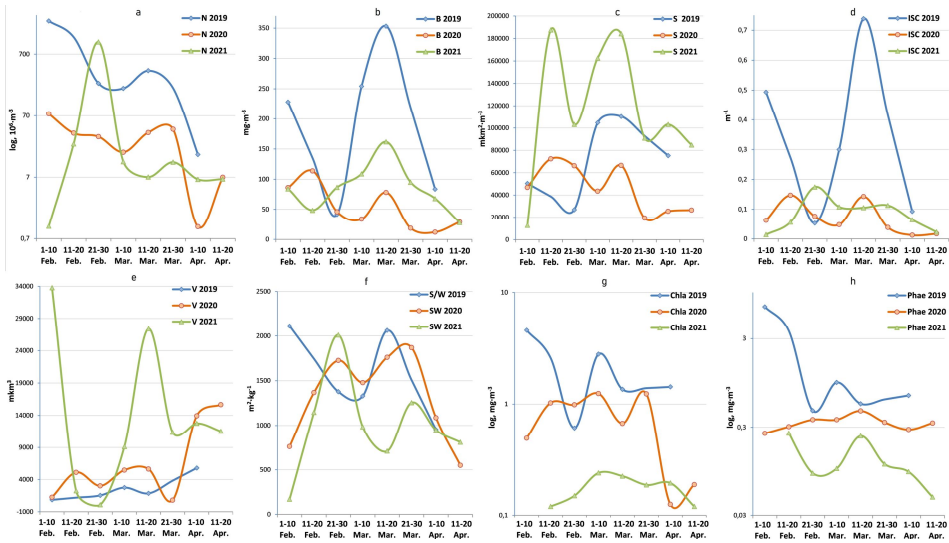


Fig. 6. Dynamics of abundance (N), biomass (B), surface area (S), surface index (ISC), average cell volume (V), specific surface area (S/W), chlorophyll *a* (Chla) and pheophytin (Phae) of phytoplankton communities in the waters of the Argentine Islands in the late summer period of 2019, 2020 and 2021 (see Supplementary, Table 6 for standard deviations).

In September 2020, a synchronous increase began, reaching the first maximum in October ($77 \text{ mg} \cdot \text{m}^{-3}$), followed by a decline in December 2020 (to $43 \text{ mg} \cdot \text{m}^{-3}$) (Fig. 7). From December to April 2021, two more successively increasing biomass maxima were observed: in Janu-

ary ($98 \text{ mg} \cdot \text{m}^{-3}$) and in March ($125 \text{ mg} \cdot \text{m}^{-3}$). Meanwhile, only one winter-spring maximum was observed in 2021 for the other development metrics: the abundance peaked in February, and the surface area and surface index peaked in March 2021 (Fig. 7).

Phytoplankton morphological structure

Morphological organization of phytoplankton in the Argentine Islands waters was characterized by a decrease in the communities' specific surface during the late summer periods of 2019, 2020 and 2021. S/W values in 2020 and 2021 were significantly lower than in 2019 (Fig. 5f; Supplementary, Table 2). This correlates with the opposite dynamics of a more traditional morphological metric – a cell volume, which was significantly bigger in 2019 than in 2020 and 2021 (Fig. 5e; Supplementary, Table 2). Phytoplankton

S/W dynamics indicates a decrease in the potential intensity of production processes in 2020 and 2021. This could be potentially driven by a decrease in the presence of small, highly functional cryptophytes and *Phaeocystis antarctica* in late summer phytoplankton in 2020 and 2021 (Figs. 3, 4). Instead, the relative contribution of large-celled diatoms to the biomass of communities has increased by almost 20% for the late summer seasons of 2020 and 2021 (Fig. 3).

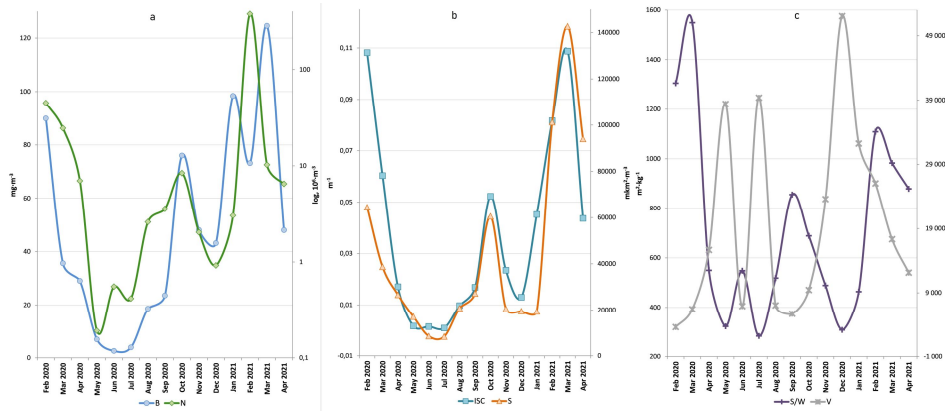


Fig. 7. Dynamics of average monthly abundance (N , $10^6 \cdot m^{-3}$), biomass (B , $mg \cdot m^{-3}$), surface area (S , $\mu m^2 \cdot m^{-3}$), surface index (ISC , m^{-1}), average cells' volume (V , m^3) and specific surface (S/W , $m^2 \cdot kg^{-1}$) of phytoplankton in the Argentine Islands waters in 2020-2021 (see Supplementary, Table 7 for standard deviations).

Morphological restructuring of phytoplankton is ensured both by the redistribution of different taxonomic groups' contributions to the communities' metrics, and by the variability of the size and shape of algae cells that form these groups. In the late summer seasons of 2020 and 2021, S/W for all mass taxonomic groups was lower than in 2019.

Indeed, the average specific surface of *Bacillariophyceae* cells was $2458 m^2 \cdot kg^{-1}$, 1471 and $2068 m^2 \cdot kg^{-1}$ in 2019, 2020 and 2021 respectively. The S/W of *Coccolithophyceae*, *Cryptophyceae* and *Dinoflagellata* decreased as well. Thus, the differences in phytoplankton morphological organization were associated with both a decrease in the presence of small-celled *Coccolithophyceae* and *Cryptophyceae* in the late summer seasons of 2020 and 2021 from 2019, as well as with a decrease in S/W of all mass groups of algae (such as *Bacillariophyceae*, *Coccolithophyceae*,

Cryptophyceae and *Dinoflagellata*).

The intra-seasonal S/W dynamics were also different in three consecutive sampling years. Two phytoplankton specific surface maxima were observed (in the third decade of February and March) in 2020 and 2021 (Fig. 6f), whereas 2019 was characterized by the peak S/W values in early February and March 2019 (Fig. 6f). A decrease in communities' S/W was observed from 1549 to $323 m^2 \cdot kg^{-1}$ during March-May 2020 (Fig. 7), followed by two similar minima, which were replaced by gradually increasing peaks of S/W from spring to summer – in June 2020 ($548 m^2 \cdot kg^{-1}$), in September 2020 ($855 m^2 \cdot kg^{-1}$) and in February 2021 ($1112 m^2 \cdot kg^{-1}$) (Fig. 7c). This confirms the morphological restructuring of Antarctic phytoplankton groups associated with the successive increase in the intensity of production processes from winter to summer, alternating with the periods of low-activity algae dominance.

Chlorophyll concentrations

Gradual and significant decrease of the average chlorophyll *a* (from 2.81 to $0.16 mg \cdot m^{-3}$) and pheophytin (from 2.82 to

$0.19 mg \cdot m^{-3}$) concentration was observed in the late summer periods of 2019, 2020 and 2021 (Fig. 5, g, h; Supplementary, Ta-

ble 2). Both chlorophyll *a* and pheophytin values correlated positively with phytoplankton biomass – $R = 0.65$, $p < 0.05$ and $R = 0.34$, $p < 0.05$ respectively. Yet, the relationship between the biomass and chlorophyll *a* concentration was different during 3 consecutive years (Fig. 8a). This may be driven by inter-seasonal differences in phytoplankton morphological or-

ganization, which, alongside with the communities' functional traits, determine the specific contribution of chlorophyll per unit of biomass. The intra-seasonal variability of pigment concentrations for late summer periods of different years had an individual character and was not characterized by common features (Fig. 6g, h).

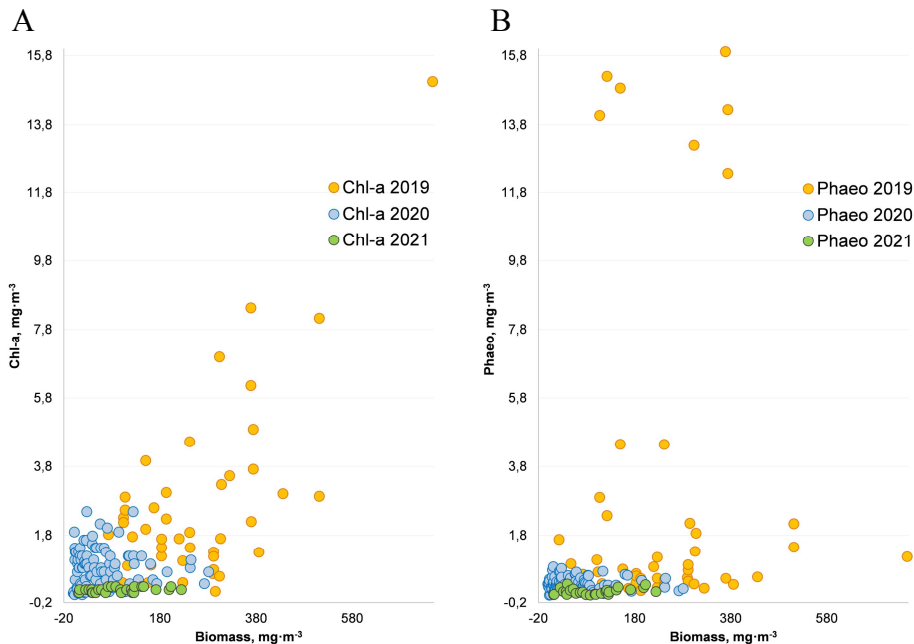


Fig. 8. Distribution of chlorophyll *a* and pheophytin concentration depending on the phytoplankton biomass in the Argentine Islands waters in the late summer period of 2019, 2020 and 2021.

Discussion

Species diversity of phytoplankton in the waters of the Argentine Islands

The relatively low phytoplankton species diversity detected in the Argentine Islands waters during the late summer of 2019 – 2021 is in line with the previous results. The list of unicellular algae species of this area was gradually supplemented starting with the first seasonal surveys in March 1997, which revealed 23 species of

diatoms (Ivanov and Minicheva 1998). 70 phytoplankton taxa were identified in this water area in the fall of 1998 (Seryogin et al. 2003), 136 taxa – in the fall of 2002 (Kuzmenko 2004), and 64 taxa – in the fall of 2004 (Gerasimiyuk 2008). Longer annual research (2002 – 2003) revealed 126 species of phytoplankton (Kuzmenko

and Ignatiev 2007, 2008). This list was supplemented in 2005 with the other 31 taxa (Kuzmenko and Ignatiev 2007, 2008). 105 and 115 species were identified in the water area of Galindez island in 2005 – 2006 and 2006 – 2007, respectively (Kuzmenko and Ignatiev 2012). The general list of phytoplankton and benthic species found in the surface water layer of the Argentine Islands archipelago contained 199 taxa (Kuzmenko and Ignatiev 2012). These estimates are close to the phytoplankton species diversity at the shelf zone of the Antarctic Peninsula, the South Shetland Islands, the Bransfield and

Drake Straits, and the ice-free waters of the Bellingshausen Sea (Kuzmenko and Ignatiev 2012, Bidigare *et al.* 2001). The diversity of phytoplankton reached 66 species in the water area of the Argentine Islands during the late summer of 2005, and the average for different seasons was 63 taxa (Kuzmenko and Ignatiev 2006, 2008). This corresponds to the average species richness detected in the late summer period of 2019 – 2021 (77 taxa). Thus, the 2019 – 2021 research did not reveal significant changes in the species diversity of the research area.

Succession of phytoplankton in the waters of the Argentine islands

Phytoplankton development cycles are determined by annual cycles of solar activity in the southern latitudes (Clarke *et al.* 2008, Ducklow *et al.* 2012). Late summer decline in the WAP phytoplankton development is followed by a winter decline in biomass with a predominance of picoplankton (Clarke *et al.* 2008, Rozema *et al.* 2017). An increase in daylight, in the angle of solar inclination and in the ice-free area accompanying spring leads to elevated light availability and activates algal growth. Ice melting contributes to the formation of a stable ocean surface layer of lower density, high temperature and high biogenic elements content. This summer stratification creates favourable conditions for the rapid algal development (Clarke *et al.* 2008, Venables *et al.* 2013, Vernet *et al.* 2008) and is considered to be a major driver of phytoplankton productivity (Garibotti *et al.* 2005a, Vernet *et al.* 2008). Indeed, macronutrients and micronutrients in the WAP coastal waters are generally considered abundant, due to the input from glaciers and upper circumpolar deep water layers (Annett *et al.* 2017, Sherrell *et al.* 2018, Carvalho *et al.* 2020).

Diatoms, cryptophytes, haptophytes, prasinophytes and mixed flagellates con-

stitute the majority of coastal phytoplankton in the WAP waters (Varela *et al.* 2002, Garibotti *et al.* 2005a, Schofield *et al.* 2017). Diatoms are a dominant taxonomic group (Mitchell and Holm-Hansen 1991, Prézelin *et al.* 2000, Schofield *et al.* 2017, Rozema *et al.* 2017). In addition, periodic blooms are caused by *Phaeocystis antarctica* (Annett *et al.* 2010) and cryptophytes, with more and more frequent mass development (Montes-Hugo *et al.* 2009, Schofield *et al.* 2017).

The first systematic seasonal phytoplankton monitoring in the waters of the Argentine Islands was conducted from April 2002 to February 2003 (Seregin *et al.* 2005, Kuzmenko and Ignatiev 2007). A gradual decrease in algal abundance (from 146.9 to 0.1 $10^6 \cdot m^{-3}$) and biomass (from 168.0 to 0.1 $mg \cdot m^{-3}$) was detected from the beginning of April to July 2002 (Fig. 9) (Kuzmenko and Ignatiev 2007). An intense *Phaeocystis pouchetii* (Hariot) *Lagerh* bloom (up to 108 $10^9 \cdot m^{-3}$ and 12.3 $g \cdot m^{-3}$) starting the beginning of the Antarctic spring (September 2002), peaking in mid-October and collapsing in November 2002 was observed. The biomass was formed by large diatoms from the *Coscinodiscus*, *Odontella*, *Membraneis*,

Thalassiora, and *Charcotia* genera during this period. In November – December, *Fragilariopsis*, *Achnanthes*, and *Corethron* dominated ($2.2 \cdot 10^9 \cdot \text{m}^{-3}$; $3.2 \text{ g} \cdot \text{m}^{-3}$) (Kuzmenko and Ignatiev 2007). Phytoplankton structure changed significantly, accompanied by a decrease of its biomass ($43.7 - 11.9 \cdot 10^6 \cdot \text{m}^{-3}$ and $93.0 - 12.1 \text{ mg} \cdot \text{m}^{-3}$) at the beginning of January 2003. Subsequently, the development of large ($330 - 650 \mu\text{m}^3$) flagellates from the *Cryptomonas* (*Cryptophyta*) and *Pyramimonas* (*Chlorophyta*) genera ($1 \cdot 10^9 \cdot \text{m}^{-3}$ and $350 \text{ mg} \cdot \text{m}^{-3}$ respectively) began in mid-January. The development of these algae was much more massive (up to $300 \cdot 10^9 \cdot \text{m}^{-3}$ and

$120 \text{ g} \cdot \text{m}^{-3}$) in the adjacent waters of Petermann Island. A prominent decrease of phytoplankton abundance and diversity was observed by the end of January (Fig. 9; Kuzmenko and Ignatiev 2007). Thus, three phytoplankton development phases, differing in abundance, biomass, diversity and size structure have been detected during the spring-summer season of 2002 – 2003: the October bloom, caused by *Phaeocystis*, was followed by the development of large diatoms in November, and by representatives of *Cryptomonas*, *Pyramimonas* and small flagellates in mid-January (Kuzmenko and Ignatiev 2007, 2008).

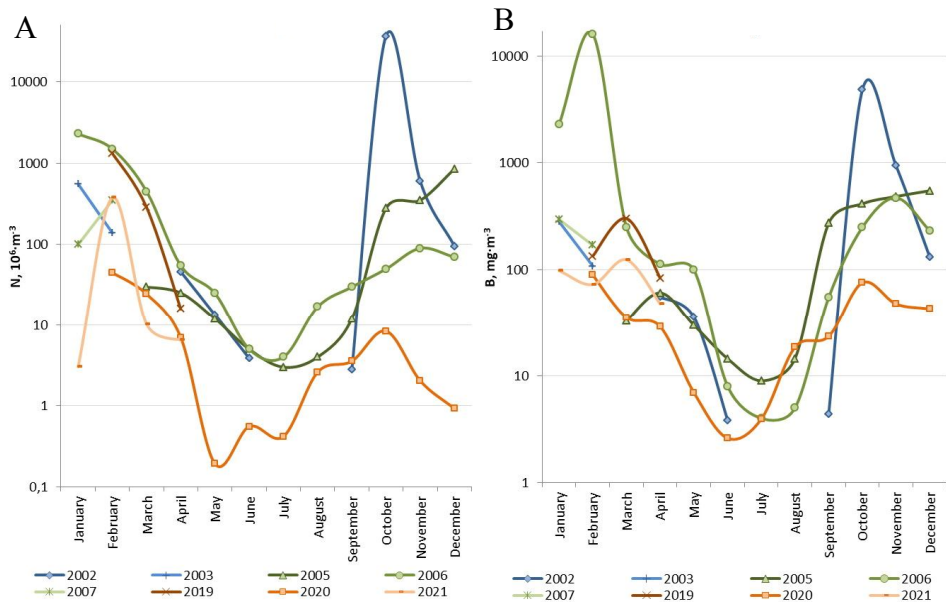


Fig. 9. Monthly phytoplankton abundance (N , $10^6 \cdot \text{m}^{-3}$) and biomass (B , $\text{mg} \cdot \text{m}^{-3}$) in the Argentine Islands waters in 2002-2003 and 2005-2007 (according to Kuzmenko and Ignatiev 2007, 2008, 2012) and 2020-2021.

A slightly different nature of phytoplankton development was witnessed in the waters of the Argentine Islands in 2005 – 2007 (Fig. 9) (Kuzmenko and Ignatiev 2012). The phytoplankton abundance and biomass gradually decreased to $1.4 \cdot 10^6 \cdot \text{m}^{-3}$ and $3.7 \text{ mg} \cdot \text{m}^{-3}$ from March to August

2005 with subsequent increase up to $214 \cdot 10^6 \cdot \text{m}^{-3}$ and $487 \text{ mg} \cdot \text{m}^{-3}$ in September – November 2005 (mainly due to the development of *Fragilariopsis* species). A bloom event was observed in summer during December 2005 – January 2006 accompanied by a peak biomass

of $2.9 \cdot 10^9 \cdot \text{m}^{-3}$ (Fig. 9). Small flagellates, as well as *Dunaliella*, *Pyramimonas* and *Cryptomonas* representatives constituted up to 61% of the abundance during this period. Meanwhile, medium and large species of diatoms (*Odontella*, *Proboscia*, *Eucampia*, *Thalassiosira* genera) accounted for up to 90% of the total biomass, and increased considerably in their biomass in response to higher water temperature, forming the most intense diatom bloom in the whole study period (up to $29.4 \text{ g} \cdot \text{m}^{-3}$; Fig. 9). The community biomass decreased to $349 \text{ mg} \cdot \text{m}^{-3}$, and the abundance (up to 85% of which was formed by *Fragilariopsis*) – to $434 \cdot 10^6 \cdot \text{m}^{-3}$ at the beginning of March 2006. Phytoplankton biomass was still mainly formed by large diatoms *Corethron*, *Trachyneis*, *Haslea*, *Membraneis* and *Proboscia* in April – May, but dropped to $14 \text{ mg} \cdot \text{m}^{-3}$, while the abundance – to $4 \cdot 10^6 \cdot \text{m}^{-3}$. No mass phytoplankton development was observed in the spring-summer season of 2006 – 2007, and the average biomass was 25 times lower than the year before, with a similar species composition (Kuzmenko and Ignatiev 2012).

Thus, in contrast to the mass development in October 2002, low occurrence of *Phaeocystis* was detected in the spring of 2005, and it was almost absent in November 2006. *Fragilariopsis* abundance increased significantly in 2005 compared to 2002 – 2003, while becoming an order of magnitude lower in the spring of 2006. The diatom *Achnanthes brevipes*, dominant in October – November 2002, was almost never found in the spring of 2005 and 2006. *Corethron criophilum*, which constituted up to 70% of biomass in autumn-winter of 2002, was also found in small quantities during 2005 – 2006 (Kuzmenko and Ignatiev 2012).

The dynamics of phytoplankton development revealed in the current study in 2019 – 2021 was comparable to previous results gained in the waters of the Argentine Islands. The quantitative development

metrics of late summer phytoplankton were found to be rather close in different periods. Thus, the average biomasses in February 2019 – 2021 are comparable to similar metrics in February 2003 and 2007, while the March biomasses of 2019 – 2021 correspond to the values observed in March 2005 and 2006. Naturally, the April biomasses of 2019 – 2021 are very close to similar values in April 2002, 2005 and 2006 (Fig. 9). The taxonomic structure of phytoplankton communities was found to be very similar during the corresponding periods of different years. Indeed, the main dominant group was large centric diatoms, with a significant contribution of smaller pennate diatoms (*Fragilariopsis*) and flagellate algae. Thus, the development of phytoplankton from February to April 2019 – 2021 corresponded to the results of previous studies, both quantitatively and qualitatively.

Nevertheless, the phytoplankton development found from February 2020 to February 2021 were lower than those in 2002 – 2007. This period was characterized by the minimum values of average monthly biomass for the entire period of phytoplankton research in the waters of the Argentine Islands. The exception was the winter-spring, as the biomass in August and September 2020 exceeded similar values of 2002, 2005, and 2006. But no bloom or even a significant phytoplankton development dominated by large centric diatoms was detected in the samples taken during the Antarctic spring of 2020. Instead, smaller *Fragilariopsis cylindrus* and *Chaetoceros* sp. with a rather low occurrence of undetermined flagellate were forming the maximum abundance during this period. Our two-week sampling periodicity leaves high probability of missing significant events in the phytoplankton development, so this atypical nature of the annual succession requires a more detailed analysis.

Hence, fairly clear patterns of phytoplankton annual succession can be detect-

ed in Argentine islands waters. After the winter decline, when the average biomass decreases to the $1\text{--}10\text{ mg}\cdot\text{m}^{-3}$, spring development begins, usually with high abundances of *Fragilariopsis* representatives and with a predominant contribution of the large centric diatoms to the biomass. At the beginning of summer, the role of *Cryptomonas* (Cryptophyta), *Pyramimonas* (Chlorophyta) and small flagellates increases. The average spring-summer biomass varies from $100\text{ to }1000\text{ mg}\cdot\text{m}^{-3}$. The dominance of large centric diatoms is established again in the late summer period, accompanied by a significant presence of small flagellates and a periodic increase in *Fragilariopsis* contribution. The average biomass gradually decreases during

the Antarctic autumn, being in the range from $10\text{ to }100\text{ mg}\cdot\text{m}^{-3}$ in April – May. A decrease in the phytoplankton development and in the role of flagellates was observed in 2020 – 2021, but the relatively low sampling frequency reduces the possibilities to interpret this data.

Only two notable exceptions have been recorded during the entire period of phytoplankton research in the Argentine Islands waters, with the single-celled algae biomass exceeding $1\text{ g}\cdot\text{m}^{-3}$. This was the spring bloom of *Phaeocystis* in October 2002 and the biomass "flash" in February 2006, formed by large centric diatoms with a significant abundance of flagellate algae.

Phytoplankton succession of the WAP area adjacent to the Argentine islands

Despite common features, the development of phytoplankton in different areas of the WAP shows significant interannual and regional variability, which can be driven by the complex effect of the regional ice conditions of the preceding winter and by the wind activity. It is hypothesized that the years with less sea ice and stronger wind mixing lead to larger input of nutrients from deeper layers to surface waters (Annett et al. 2010). On the other hand, less deep water column mixing driven by moderate wind activity and high sea ice extent requires less solar energy to stabilize it in the following summer (Venables et al. 2013). Furthermore, more algal biomass can accumulate in more stratified waters, leading to phytoplankton blooms (van Leeuwe et al. 2020). Recent climate change consequences are also affecting phytoplankton succession in the WAP waters (van Leeuwe et al. 2020). Comparison of the phytoplankton dynamics in the Argentine Islands waters in 2019 – 2021 with the processes detected in the adjacent areas of the WAP in previous years can contribute to understanding of

the regional trends.

Research conducted south of the Argentine Islands, at Rothera Station in Ryder Bay in 2012 – 2017 revealed clear patterns of phytoplankton succession (van Leeuwe et al. 2020). Two consecutive blooms were observed for four out of the total five observation years, following previously described patterns (Clarke et al. 2008, Annett et al. 2010). Multi-year data confirmed a clear species succession between November and March. Autotrophic flagellates (haptophytes, chlorophytes, and prasinophytes), most likely derived from sea ice communities, were dominant in spring and gradually declined by January. *Chrysochromulina* sp., *Micromonas* sp., *Pyramimonas* spp. and *Phaeocystis antarctica* were dominating. Diatoms peaked in more stratified and warmer surface waters from January to March. In January – February, small centric diatoms *Thalassiosira* sp. ($< 20\text{ }\mu\text{m}$) were quite common, but such large centric species as *Proboscia inermis* and *Corethron* sp. formed the major biomass. In January, a significant abundance of small cryptophytes ($5\text{--}20\text{ }\mu\text{m}$) with strong signs

of mixotrophy were also observed, whereas dinoflagellates were present in very low abundance (van Leeuwe *et al.* 2020). However, significant interannual fluctuations associated with differences in wind mixing, ice cover, and El Niño were detected, which began in December 2015 and had a strong impact on processes in the WAP (van Leeuwe *et al.* 2020). High chlorophyll content was observed in 2013 and 2017 under conditions of a shallow summer mixed layer (van Leeuwe *et al.* 2020). The intervening 2014 – 2016 years were characterized by low biomass, which contradicts the hypothesis of the link between heavy winter sea ice cover and subsequent high summer phytoplankton biomass (Venables *et al.* 2013).

Remarkably similar annual phytoplankton successional patterns driven by constant seasonal factors (*e.g.*, solar radiation, temperature, and freshwater input) were shown at the Palmer Station, located north of the Argentine Islands (2017 – 2019), while storm/wind events caused more ephemeral interannual differences (Nardelli *et al.* 2023). A shorter sea ice season with early rapid ice retreat was associated with low phytoplankton biomass and reduced diatom contributions (2017 – 2018), whereas a longer sea ice season with late ice retreat resulted in the opposite (2018 – 2019) (Nardelli *et al.* 2023). Consistent seasonal succession patterns were detected despite significant interannual differences in sea ice dynamics and phytoplankton biomass (Nardelli *et al.* 2023). They correspond to earlier studies in the coastal WAP (Schofield *et al.* 2017, Garibotti *et al.* 2005a), but differ from those found in the Rothera station area (van Leeuwe *et al.* 2020).

Three main successive phases of phytoplankton development were defined for the Palmer Station area: spring ice retreat phase, peak summer phase and late summer phase. The spring ice retreat phase (November–December) was characterized by blooms dominated by large centric

diatoms associated with the spring sea ice retreat and upper water layer stratification. This phase was marked by a high abundance of centric diatoms (55% of the total biovolume), prasinophytes *Pyramimonas spp.* (5% of biovolume) and *P. antarctica* haptophytes (3% of biovolume) in 2017 – 2019 (Nardelli *et al.* 2023). Large spring blooms of centric diatoms have been observed in other Antarctic studies (Garibotti *et al.* 2005 a,b; Costa *et al.* 2020, 2021). Yet, these events are significantly different from the succession at the Rothera station, where the initial phase starts with the flagellates – *Pyramimonas spp.* and *Phaeocystis antarctica*, and is fuelled by the sea ice melt (van Leeuwe *et al.* 2020, 2022). These species were also observed in relatively high abundance at the Palmer station during the spring ice retreat phase, but their relative dominance was not pronounced. There was a rapid increase in centric diatoms such as *Thalassiosira spp.* and *Chaetoceros spp.*, which are absent in sea ice and gain advantage from the optimal light conditions provided by stratification (Nardelli *et al.* 2023).

The peak summer phase (December–January), accompanied by reduced nutrient reserves, is characterized by the prevalence of mixotrophic cryptophytes with low chlorophyll *a* content. The lowest chlorophyll *a* concentrations, yet the highest biovolume and abundance values, were recorded during the peak summer phase of 2017 – 2019. It was associated with abundant cryptophytes (29% biovolume), mixed flagellates (27% biovolume) and unidentified small-celled pennate diatoms, probably *Fragilariopsis spp.* and *Nitzschia spp.* (19% biovolume) (Nardelli *et al.* 2023). The vigorous development of these pennate diatoms produced almost twice as much biomass as any other bloom during the study period (Nardelli *et al.* 2023).

During the late summer phase (February–March), small (< 20 µm) diatoms and mixed flagellates, including haptophytes and undetermined flagellates, dominated

(Nardelli et al. 2023). The late summer phase was associated with a relatively high abundance of centric diatoms (47% of bio-volume) and mixed flagellates (30% of biovolume). 46% of the centric diatoms were unidentified disc-shaped cells 15 to 20 μm in diameter (probably of the genus *Thalassiosira*). Species of pennate diatoms, including *Cocconeis* spp. and *Licmophora* spp., had the highest abundance during this phase (Nardelli et al. 2023).

In summary, common features of phytoplankton development can be detected along the WAP. The October 2002 *Phaeocystis* bloom in Argentine Islands waters corresponds to the first stage detected at the Rothera station, while in all subsequent

years the spring-summer development of phytoplankton was close to that described in the studies at the Palmer station. However, the late summer phase at the Palmer station area is almost not manifested in the waters of the Argentine Islands, except for 2019, when *Phaeocystis antarctica*, cryptophyte *Hillea fusiformis* and *Cryptomonas* sp. developed at the end of summer. The development of phytoplankton during this period corresponded to the stage of succession, more typical for southern Ryder Bay. Further monitoring has a potential to reveal changes occurring in the organization of marine phytoplankton communities under the impact of climate change.

Conclusion

121 species and intraspecific taxa of unicellular algae belonging to 12 taxonomic classes were found in the waters of the Argentine Islands during the late summer of 2019 – 2021. *Bacillariophyceae* completely dominated the studied phytoplankton community. The biomass dynamics was largely determined by the development of the most massive diatom *Corethron criophilum* and a complex of other large-celled centric diatoms with low specific surface values ($70 - 300 \text{ m}^2 \cdot \text{kg}^{-1}$). The nature of late summer succession was typical for the Argentine Islands waters in 2019, 2020 and 2021. However, in contrast to 2020 and 2021, a more significant development of small-celled species with specific surface values exceeding $700 \text{ m}^2 \cdot \text{kg}^{-1}$ (haptophyte *Phaeocystis antarctica* ($785 \text{ m}^2 \cdot \text{kg}^{-1}$), cryptophytic *Hillea fusiformis* ($775 \text{ m}^2 \cdot \text{kg}^{-1}$) and *Cryptomonas* sp. ($1011 \text{ m}^2 \cdot \text{kg}^{-1}$), as well as small-celled pennate diatom *Fragilariopsis cylindrus* ($2371 \text{ m}^2 \cdot \text{kg}^{-1}$)) was observed in 2019. The lowest metrics of phytoplankton development along with the simplified community structure were detected during April 2020 – February 2021. Specific surface values

were significantly lower in late summer 2020 and 2021 than in 2019 indicating a lower intensity of functional processes.

Clear annual succession patterns have been detected in Argentine Islands' phytoplankton communities. The winter recession with the average biomass drop to $1-10 \text{ mg} \cdot \text{m}^{-3}$ was followed by spring development, associated with high *Fragilariopsis* abundance and predominant contribution of large centric diatoms. The role of a *Cryptomonas* (Cryptophyta), *Pyramimonas* (Chlorophyta) and small flagellates increased at the beginning of summer. Subsequently, the dominance of large centric diatoms with a significant presence of small flagellates and a periodic increase in the contribution of *Fragilariopsis* was established in late summer period.

Phytoplankton development observed in Argentine Islands' waters resembled the processes detected in different geographically close regions along the WAP, namely the Rothera and Palmer stations, yet was characterized by certain peculiarities. Further observations might shed the light on the impact of climate change on marine phytoplankton communities.

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Supplementary Materials

The below-specified Tables are available and freely downloadable at the CPR web page (Vol. 14, No. 2, Zotov et al. 2024).

Table 1. Coordinates of phytoplankton monitoring stations in the waters of the Argentine Islands.

Table 2. Interannual differences in average abundance (N), biomass (B), surface area (S), surface index (ISC), average cell volume (V), specific surface (S/W), chlorophyll a concentration (Chl a) and pheophytin (Phae) of phytoplankton groups in the waters of the Argentine Islands in the late summer period of 2019-2021 (Multiple Range Tests for 95.0% LSD).

Table 3. Distribution of phytoplankton abundance ($10^6 \cdot m^{-3}$) between taxonomic classes in the Argentine Islands waters in late summer period 2019-2021.

Table 4. Distribution of phytoplankton biomass ($mg \cdot m^{-3}$) between taxonomic classes in the Argentine Islands waters in late summer period 2019-2021.

Table 5. Distribution of phytoplankton surface area ($m^2 \cdot m^{-3}$) between taxonomic classes in the Argentine Islands waters in late summer period 2019-2021.

Table 6. Averages and standard deviations for abundance (N, $10^6 \cdot m^{-3}$), biomass (B, $mg \cdot m^{-3}$), surface area (S, $\mu m^2 \cdot m^{-3}$), surface index (ISC, m^{-1}), specific surface (S/W, $m^2 \cdot kg^{-1}$) and average cells' volume (V, m^3) of phytoplankton in the Argentine Islands waters in 2019-2021.

Table 7. Monthly averages and standard deviations for abundance (N, $10^6 \cdot m^{-3}$), biomass (B, $mg \cdot m^{-3}$), surface area (S, $\mu m^2 \cdot m^{-3}$), surface index (ISC, m^{-1}), specific surface (S/W, $m^2 \cdot kg^{-1}$) and average cells' volume (V, m^3) of phytoplankton in the Argentine Islands waters in 2020-2021.